

1

2 Social Perception and Supply of Ecosystem Services – A 3 Watershed Approach for Carbon Related Ecosystem 4 Services

5 Antonio J. Castro, Caryn C. Vaughn, Jason P. Julian,
6 Marina García Llorente and Kelsey N. Bowman

7 Additional information is available at the end of the chapter

9 1. Introduction

10 Over the past two decades research on ecosystem services, i.e. the benefits that humans derive
11 from natural systems [1], has gained importance among scientists, managers, and policy-
12 makers worldwide as a way to communicate societal dependence on ecological life support
13 systems integrating both natural and social science perspectives [2]. Ecosystem services can
14 be direct benefits, such as food or freshwater for drinking, or indirect benefits through
15 provisioning of services such as carbon sequestration [1]. Ecosystem services include 1)
16 provisioning services obtained directly from the ecosystem such as food provision, 2) regu-
17 lating services such as water regulation, habitat, air quality, and water quality, and 3) cultural
18 services, which are the benefits that people obtain through tourism, aesthetic values, spiritual
19 enrichment, and sense of place [3, 4].

20 The ecosystem services approach is useful for decision-making in conservation and natural
21 resource management [5] because it assigns value to nature by translating ecosystem proper-
22 ties into human needs [6]. Ecosystem services can be valued using different approaches
23 ranging from biophysical quantifications to sociocultural surveys to economic assessment.
24 Biophysical quantification of services such as carbon storage and sequestration have recently
25 been used extensively in conservation applications. However, to conserve biodiversity, we
26 need to move beyond narrow studies of species or habitat status and increase social awareness
27 of the broader importance of conservation [2]. A key challenge in implementing this approach
28 is identifying an ecosystem's capacity to provide services (supply side) and the social demand
29 for those services (demand side). Addressing both the supply and demand for ecosystem

1 services underscores the fact that the importance of an ecosystem service to people is influ-
2 enced not only by the ecosystem's properties but also by societies need for that service and
3 how that need is perceived.

4 The Kiamichi River watershed, in southeastern Oklahoma (USA), provides many direct and
5 indirect ecosystem services to stakeholders that live in or visit the area. This watershed and
6 the area surrounding it is a national biodiversity hotspot, meaning it is biologically rich, yet
7 threatened. This area is also at the center of a highly politicized debate between different
8 stakeholders' plans for the use of the watershed's ecosystem services and activities that may
9 affect those services [7]. The Kiamichi watershed not only provides many important freshwater
10 services (e.g., drinking water, water filtration or recreation), but it provides numerous
11 terrestrial ecosystem services as well such as habitat for species and food production. The land
12 is relatively undeveloped with few urban areas and extensive tracts of second growth, forested
13 landscapes [8] that provide carbon storage and sequestration. Carbon sequestration is
14 considered an optimal descriptor of ecosystem functioning [9, 10, 11]. It is a current focus in
15 climate change studies and is classified as an intermediate service [12] or as supporting the
16 delivery of other regulating services [13]. Most people are unaware that carbon sequestration
17 provides direct benefits such as erosion control and soil fertility and indirect benefits such as
18 air quality and habitat for species.

19 Here, we used the Kiamichi River watershed as a case study to examine the social perception
20 and biophysical supply of carbon related services. We first assessed the social perception of
21 the general public regarding a variety of ecosystem services provided by the Kiamichi
22 watershed in southeastern Oklahoma, including direct and indirect benefits related to the
23 carbon cycle. We used a carbon sequestration model to quantify the spatial distribution of
24 carbon storage and sequestration across the watershed. We used these results along with the
25 social perception of services and the watershed capacity for carbon sequestration to analyze
26 the supply-demand framework of ecosystem services [14]. Finally, we discuss the implications
27 for linking the structure and functioning of biodiversity within the watershed.

28 **2. Problem statement**

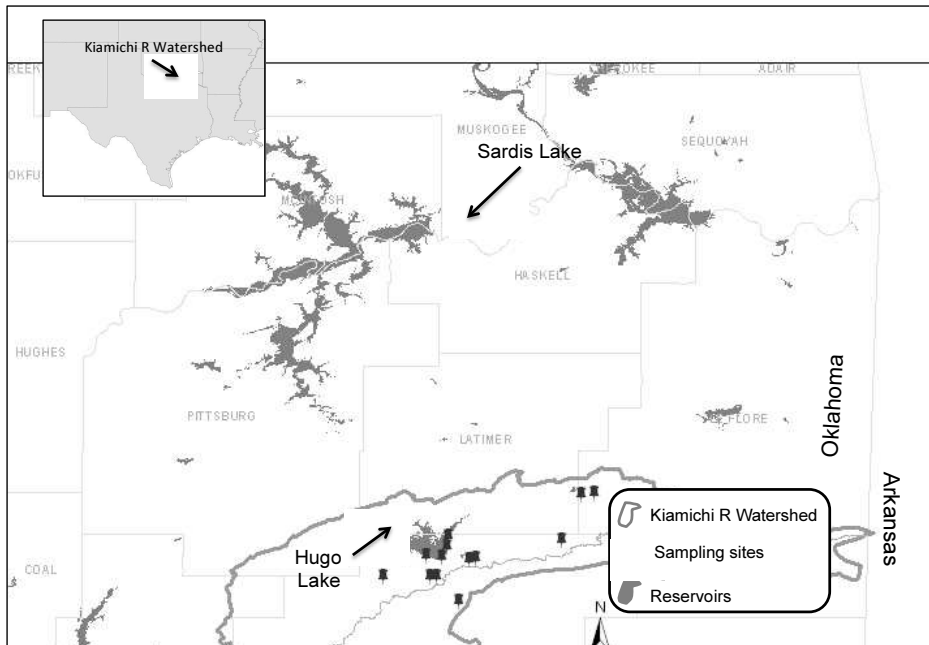
29 Changes in land use-land cover are recognized as one of the most important direct drivers in
30 ecosystem services delivery [15]. Landscapes across the U.S. are changing with human
31 population growth and increased development. These land use changes alter the natural sinks
32 and pools of carbon in the environment, but are often not included in land management or
33 planning. Different land use types and dominant vegetation differ in their storage capacity
34 and sequestration rate [15]. To better understand the impacts of land cover-use changes in
35 relatively undeveloped areas such as the Kiamichi watershed, research is needed on different
36 land uses and land changes and their impacts on carbon storage and sequestration.

37 Carbon sequestration can be viewed as an optimal descriptor of ecosystem functioning [10,11],
38 and human-derived carbon fluctuations [16] in the atmosphere affect many other services such
39 as air quality and biomass production. Changes in air quality are one of the carbon related

1 ecosystem services that is most easily recognized by the public. Thus, understanding how the
 2 public perceives the status and importance of air quality can help inform resource manage-
 3 ment. Our study compares perceptions of Kiamichi watershed stakeholders with the actual
 4 state of carbon sequestration services and land use practices in the watershed.

5 3. Application area

6 The Kiamichi River watershed in southeastern Oklahoma, with a drainage area of 4,650 km²,
 7 is a major tributary of the Red River, which flows into the Mississippi River and Gulf of Mexico
 8 (Figure 1). The watershed is 64% forest, 18% pasture, 11% grassland/shrubland, 3% urban, 3%
 9 open water, and 1% wetlands according to the 2006 National Land Cover Dataset. While most
 10 of the watershed is temperate deciduous forest (primarily oak-hickory), there are several
 11 conifer plantation forests across the watershed. Its steep and rugged terrain has limited major
 12 row crop agriculture, there are no nearby major cities or interstates, and human population
 13 density is low [5.6 people / km²] [17] This lack of development in the watershed has left the
 14 Kiamichi River with relatively pristine water and high aquatic biodiversity, containing 86 fish
 15 species and 31 mussel species, three of which are federally endangered [18,19,17, 20]



16
 17 **Figure 1.** Kiamichi River watershed study area and sampling sites.

1 **4. Method used**

2 **4.1. Social sampling and analysis of perceptions of ecosystem services**

3 We conducted social sampling regarding public perceptions of a suite of ecosystem services
4 provided by the Kiamichi watershed. In summer 2013, we conducted 304 random, individual,
5 face-to-face surveys across the watershed. Interviewees included stakeholders residing in the
6 watershed, tourists, and people working within the watershed. Sampling was conducted at 30
7 sites in the watershed (Figure 1). Social preferences regarding the variety of ecosystem services
8 provided the Kiamichi River were explored through ranking [21]. Our study included eight
9 categories of carbon-and water-related ecosystem services in three classes: provisioning
10 (freshwater provision), regulating (water regulation, water quality, air quality, and habitat for
11 species), and cultural services (recreation, cultural heritage, and local identity). We asked
12 interviewees if they felt that the Kiamichi River provided benefits that contribute to human
13 well being (very much, much, not very much, and nothing), and asked them to provide
14 examples of potential benefits. All respondents were asked to indicate the relative importance
15 and perceived trend of each service over the last 10 years. To do this, they were asked to select
16 the four services most important to them and to rank them from 1 to 4 (important to essential
17 services). From this information, we created an ordinal measure of the importance of each
18 service to each respondent [22].

19 **4.2. Mapping the distribution of carbon storage and sequestration**

20 To model carbon storage we used InVEST (Integrated Valuation of Environmental Services
21 and Tradeoffs). InVEST is a family of GIS tools designed by the Natural Capital Project to
22 inform decisions about natural resource management and provides an effective tool for
23 evaluating trade-offs among ecosystem services under different scenarios [23]. InVEST models
24 are spatially explicit and return results in either biophysical (e.g., tons of carbon stored) or
25 economic terms (e.g., net present value of that sequestered carbon). We used the InVEST carbon
26 sequestration model to quantify and map the current (i.e., 2006) spatial distribution of carbon
27 sequestration across the Kiamichi watershed. Here, the carbon model estimates for each pixel
28 (30-meter resolution) a value that represents the change in storage between two time periods.
29 Negative values represent a loss in carbon sequestering capacity, and positive values represent
30 areas that have gained more capacity to sequester carbon.

31 We used InVEST Terrestrial Toolboxes (version 2.5.6) in ArcMap (10.2) to generate a map
32 of the balance of carbon sequestration in the Kiamichi watershed. The model needs several
33 inputs to successfully estimate carbon sequestration including land use-land cover (LULC)
34 maps for the two years of comparison and data on each LULC's capacity to stock carbon
35 in four fundamental carbon pools: above ground biomass, below ground biomass, soil, and
36 dead matter. These data can be collected from real time monitoring of carbon levels or from
37 the literature. We obtained carbon pool values from the 2006 IPPC Guidelines for National
38 Greenhouse Gas Inventories report by the Intergovernmental Panel on Climate Change

1 [24]. According to this source, southeastern Oklahoma is considered a subtropical steppe
2 climate. Estimated carbon values for LULC types for a subtropical steppe climate were
3 derived from various IPCC tables in Volume 4 of the report. For the five LULC types
4 selected we calculated the mean value when multiple values were available. Not all four
5 of the required carbon pools were listed for each LULC category in the IPCC report; so
6 additional literature searches were conducted [25,26]. Finally, all carbon pool values were
7 converted into metric tons (or Mega grams) per hectare (Mg ha^{-1}) and formatted in a table,
8 as per InVEST model requirements.

9 **4.3. Land use-Land Cover (LULC) maps**

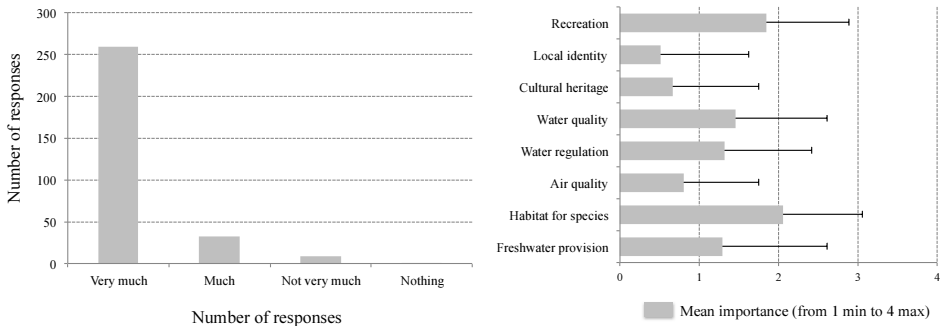
10 We compared changes in LULC between 1898 and 2006. The LULC map for 1898 is the earliest
11 complete data set for the Kiamichi watershed and served as the reference year for the carbon
12 model. LULC in 1898 largely represents the potential natural vegetation and pre-European
13 landscape of southeastern Oklahoma [27]. We created the 1898 LULC map using data from
14 [28], which was derived from Public Land Survey System records made available by the Bureau
15 of Land Management's General Land Office. Our 1898 map included four LULC categories:
16 cropland, forest, grassland, and wetland. The 2006 LULC map was derived from the National
17 Land Cover Database [29], which contained over twenty LULC categories. To make the two
18 datasets compatible with each other and InVEST, we grouped LULC as follows: Urban-Barren,
19 Water-Wetland, Forest, Shrub-Grassland-Pasture, and Cropland. The 1898 dataset was
20 converted to a 30-meter raster to match the 2006 NLCD.

21 **5. Status and results**

22 **5.1. Social perception of watershed services**

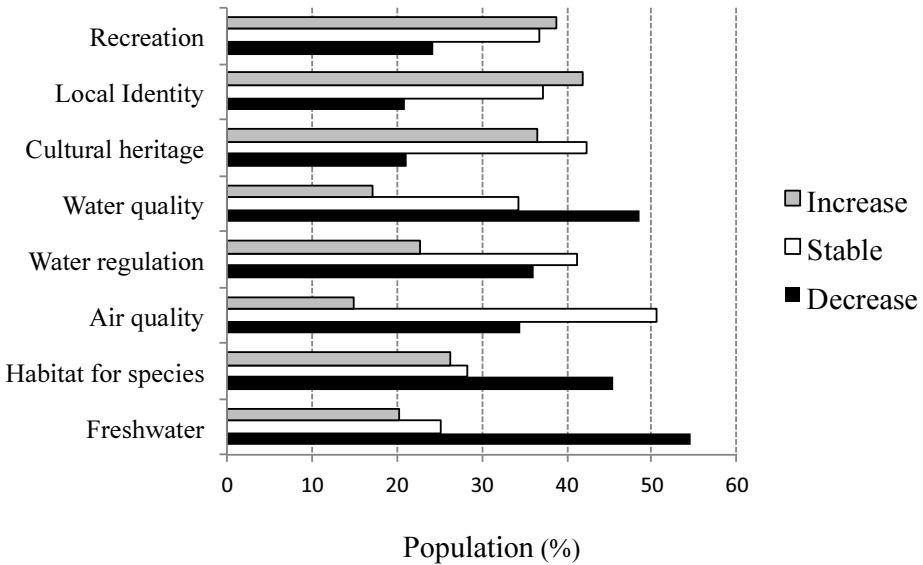
23 Of the 304 respondents, 300 (99%) answered that the Kiamichi River is "providing benefits that
24 are contributing to human wellbeing," with 86% answering that it provides substantial benefits
25 (i.e., very much, Figure 2a). Only one respondent said that no benefits were provided by the
26 Kiamichi, and three respondents did not answer the question. When asked to give an example
27 of a benefit provided by the Kiamichi, virtually all of those who responded gave an example
28 related to water resources (i.e., drinking water, fishing, recreation). Air quality was not
29 mentioned by any of the respondents as a watershed benefit.

30 The ecosystem service with the highest average importance among all respondents was habitat
31 for species, followed by recreation and water quality (Figure 2b). Ecosystem services consid-
32 ered less important were local identity, followed by cultural heritage and air quality. Most
33 respondents thought that many of the services they considered most important to human
34 wellbeing (habitat for species and water quality) had declined, while those services that were
35 not considered as important (cultural heritage and local identity) had remained stable or
36 increased (Figure 3). Air quality was considered to be the most stable ecosystem service.



65

66 **Figure 2.** Perception of Kiamichi watershed benefits and social importance of supplied ecosystem services



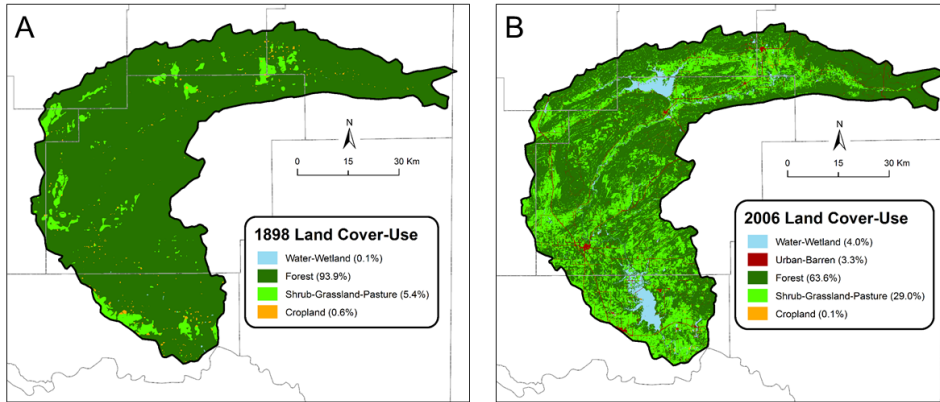
1

2 **Figure 3.** Ecosystem services trends in the Kiamichi Watershed.

3 **5.2. Land use-Land Cover (LULC) change between 1898 and 2006**

4 Changes in LULC between 1898 and 2006 are important to understanding the carbon seques-
 5 tration balance in the Kiamichi watershed. To run the sequestration model, LULC datasets for
 6 1898 and 2006 were reclassified into five categories: urban-barren, cropland, forest, shrub-
 7 grassland-pasture, and water-wetland (Figure 4). In 1898, 93.9% of the Kiamichi watershed
 8 was covered in forest, and 5.4% was covered in shrubland, grassland, and pasture. Only a fraction
 9 of a percent of the land was covered by cropland (0.6%) or water-wetland (0.1%). In 2006, the
 10 Kiamichi watershed represented a rural landscape, with many of the forests replaced with

1 pastures. The 30.3% decline in forest was largely accounted for by the 23.6% increase in shrub-
 2 grassland-pasture. Urban development (via 4 small towns) and water reservoir creation (via
 3 two large dams) accounted for the rest of the lost forests. Between 1898 and 2006, cropland
 4 decreased from 0.6% to 0.1%.



5
 6 **Figure 4.** Land cover-use maps for the Kiamichi River watershed in southeastern Oklahoma for 1898 (A) and 2006 (B).
 7 Oklahoma county boundaries are included for reference.

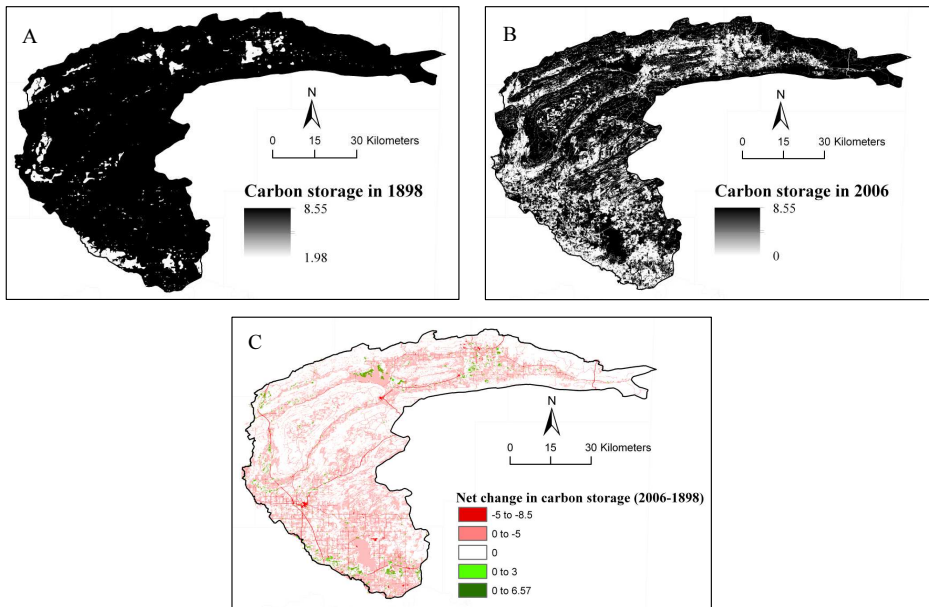
8 **5.3. Carbon storage and sequestration in the Kiamichi watershed**

9 Based on carbon stocks data collected for each LULC type (Table 1), we observed a significant
 10 decrease between 1898 and 2006 in the capacity of the watershed to store carbon (Figure 5a-
 11 b). Because of this, there are areas, mostly in the upper and lower basin, where the conversion
 12 of forests to shrub-grassland-pastures produced a decrease of total carbon storage (Figure
 13 5B). This involved a reduction in the aboveground and soil stocks, producing a difference of
 14 over 30 Mg ha⁻¹ of carbon per hectare in both the aboveground and soil carbon stocks.

Land use-land cover	Above ground biomass	Below ground biomass	Soil	Dead matter
Cropland	1.67	4.52	17.80	0.00
Forest	37.60	7.52	48.50	3.45
Shrub-Grassland-Pasture	1.27	5.08	24.05	0.13
Water-Wetland	0.00	0.00	68.25	0.15
Urban-Barren	0.00	0.00	0.00	0.00

15 **Table 1.** Carbon stock related to the four carbon pools required for InVEST carbon model. Data are converted into
 16 metric tons of carbon per hectare (Mg ha⁻¹)

1 The results from the carbon model output clearly show an overall trend of decreasing and null
 2 carbon sequestration in most of the Kiamichi watershed (Figure 5c). Considering the positive
 3 and negative carbon sequestration estimations, the carbon model obtains a watershed total
 4 of -9.197.087 metric tons of carbon. This result shows that the watershed stored 9.1 million less
 5 metric tons of carbon in 2006 than it did in 1898. There are small patches of positive carbon
 6 sequestration (green area in Figure 5c) due to recent reforestations around Sardis and Hugo
 7 reservoirs. The areas experiencing the most negative carbon sequestration (red area in Figure
 8 5c) are those areas converted from forest and grasslands into urban-barren land. The lower
 9 watershed area has experienced the most loss of carbon sequestration capacity. One explanation
 10 for this pattern of agricultural land conversion is that the lower watershed is flatter and
 11 more suitable for pasture while the steeper slopes of the upper watershed limit pasture
 12 development. However, there is still a loss in sequestration as the forested mountain slopes
 13 are being thinned for timber production.



14

15 **Figure 5.** Maps of carbon storage in 1898 (A) and 2006 (B), and net change in carbon storage during this period (C) for
 16 the Kiamichi River watershed. Positive values indicate a net-gain in carbon sequestration (e.g., cropland to forest),
 17 whereas negative values indicate lost carbon sequestration (e.g., forest to pasture). The values are in Mg/km².

18 6. Conclusions

19 Conserving ecological processes is necessary to maintain human wellbeing. The ecosystems
 20 services approach allows for quantification of the importance of ecological processes to

1 humans. Such quantification should include multiple dimensions including biophysical, socio-
2 cultural and economic valuations. Our study provides a multidimensional valuation of carbon-
3 and water-related ecosystem services in a large rural watershed. Carbon sequestration is an
4 optimal ecosystem service because it ensures the supply of other ecosystem services such as
5 food production, green areas for recreation and better air quality [13,11]. Our results show that
6 people living in the watershed think the area provides ecosystem services, but that air quality
7 is not as important as services such as habitat for species, water quality, and recreation.
8 Ecosystem services associated with water resources are highly visible (i.e. water availability,
9 recreation on lakes) and these are the most highly valued by our survey respondents. Unlike
10 water related services, air quality is less tangible and difficult for people to visualize in areas
11 without heavy air pollution. However, changes in carbon storage (the watershed lost the
12 capacity to store and sequester 9.1 million metric tons of carbon since 1898) reflect conversion
13 of natural forests into agricultural lands or timber production stands. Stakeholders in the
14 watershed need to understand that in the long term, continuing this land conversion trend
15 will decrease carbon sequestration and potentially air quality. We think our novel, multidimensional
16 approach combining both biophysical supply and social perception of carbon
17 related ecosystem services will help stakeholders and managers make more informed land use
18 decisions in the future.

19 For future research, as climate change and human development continue to interact and affect
20 the delivery of ecosystem services, other valuation practices including mapping the biophysical
21 supply of other ecosystem services such as biodiversity conservation or water regulation
22 will provide enough practical results for landscape management and planning. Currently,
23 other mapping tools such as the Artificial Intelligence for Ecosystem Services (ARIES) [30] or
24 POLYSCAPE [31] are applied to landscapes of all sizes and are expected to work well with
25 each unique scenario [2]. Many researchers in the field of biology, ecology, and environmental
26 studies are calling for a focus on multidimensional approaches that include both a natural
27 valuation component along with a social one [28].

28 **7. Study limitations, assumptions, and future work**

29 There were limitations to this study and assumptions made for the InVEST model. Regrouping
30 and simplifying LULC classes obviously generalized carbon storage losses/gains. When
31 reclassifying the 2006 LULC map, some reclassifications were obvious by the descriptions, but
32 some others required assumptions. For example, those LULC types classified as Central Oak-
33 Hardwood and Pine Forest by the National Vegetation Classification were reclassified into
34 simply Forest. Reclassification of other National Vegetation Classification LULCs, such as
35 Recently Disturbed or Modified were assumed, and requires further investigations.

36 Carbon pool data collection also presented some challenges. Because the available IPCC carbon
37 data values were based on broadly generalized values for each climate division, many
38 assumptions were made as to vegetation types in the area. In this sense, further research for
39 carbon pools for each dominant vegetation species per LULC to obtain a value that is more
40 indicative of the watershed itself, not just the climate region.

1 Further, this study only looked at two years to derive carbon storage estimates. Southeastern
2 Oklahoma is a dynamic landscape that can change at monthly and annual timescales due to
3 timber harvesting, fire, drought, and insect infestations [32]. Some studies have characterized
4 this region as having one of the highest annual rates of land cover change in the U.S. [33] and
5 as being one of the most sensitive to climate change [34]. If we want to capture these land cover
6 changes at higher spatio-temporal resolutions, new techniques will be needed [e.g., 32, 35].
7 These frequent and intense changes to forest cover have many implications for carbon storage
8 dynamics, which was also beyond the scope of our study.

9 **Acknowledgements**

10 We thank all of the people in the Kiamichi watershed that kindly responded to the question-
11 naire. We thank Melanie Lawson and Joseph Sardasti for assisting in fieldwork, and Bruce
12 Hoagland and Todd Fagin for assisting in land use-land cover maps. The Oklahoma Biological
13 Survey and the South Central Climate Science Center (SC-SCC) at the University of Oklahoma
14 (US) provided funding for the development of this research.

15 **Author details**

16 Antonio J. Castro^{1*}, Caryn C. Vaughn¹, Jason P. Julian², Marina García Llorente³ and
17 Kelsey N. Bowman⁴

18 *Address all correspondence to: acastro@ou.edu

19 1 Oklahoma Biological Survey, Department of Biology and Ecology and Evolutionary Biology
20 Graduate Program, University of Oklahoma, Norman, OK, USA

21 2 Department of Geography, Texas State University, San Marcos, TX, USA

22 3 Sociology of Climate Change and Sustainable Development research group, Department
23 of Social Analysis, University Carlos III, Madrid, Spain

24 4 Department of Geography and Environmental Sustainability, University of Oklahoma,
25 USA

26 **References**

27 [1] Daily G, Alexander S, Ehrlich, P, Goulder L, Lubchenco J, Matson P, Woodwell G.
28 Ecosystem Services: Benefits Supplied to Human Societies by Natural Ecosystems. *Is-*
29 *ssues in Ecology*; 1997.

- 1 [2] Castro A, Garcia-Llorente M, Martin-Lopez B, Palomo I, Iniesta_Arandia I. Multidi-
2 mensional approaches in ecosystem services assessment. In: Earth Observation of
3 Ecosystem Services, 2013a. 441-468.
- 4 [3] Perrings C, Naeem S, Ahrestani, FS, Bunker E, Burkill P, Canziani G, Elmqvist T,
5 Fuhrman JA, Jaksic FM, Kawabata Z, Kinzig A, Mace GM, Mooney HM, Prieur-Ri-
6 chard AH, Tschirhart J, Weisser A. Ecosystem services, targets, and indicators for
7 the conservation and sustainable use of biodiversity. *Frontiers in Ecology and the En-
8 vironment* 2011;9:512-520
- 9 [4] Wainger L, Mazzotta M. Realizing the potential of ecosystem services: A framework
10 for relating ecological changes to economic benefits *Environmental Management*
11 2011;48:710-733.
- 12 [5] Harrison PA. Ecosystem services and biodiversity conservation: an introduction to
13 the RUBICODE project. *Biodiversity Conservation* 2010;19:2767-2772
- 14 [6] Castro AJ, Martín-López B, García-Llorente M, Aguilera PA, López E, Cabello J. So-
15 cial preferences regarding the delivery of ecosystem services in a semiarid Mediter-
16 ranean región. *Journal of Arid Environment* 2011;75:1201–1208
- 17 [7] Oklahoma Water Resources Board vs. Choctaw and Chickasaw Nations of Oklaho-
18 ma. Oklahoma Supreme Court, case 110375. 2012.
- 19 [8] Mast M, Turk J. Environmental characteristics and water quality of Hydrologic
20 Benchmark Network stations in the West-Central United States, 1963-95. U.S. Geo-
21 logical Survey Circular. 1999.
- 22 [9] Castro AJ, Paruelo JM, Alcaraz-Segura D, Cabello J, Oyarzabal M, López-Carrique E.
23 Missing gaps in the estimation of the carbon gains service from Light Use Efficiency
24 models In: Alcaraz-Segura, Di Bella, C., D. (Eds). *Earth Observation of Ecosystem
25 Services*, 2013b.105-124
- 26 [10] Paruelo JM, Piñeiro G, Baldi G, Baeza S, Lezama F, Altesor AI, Oesterheld M. Carbon
27 Stocks and Fluxes in Rangelands of the Río de la Plata Basin. *Rangeland Ecosystem
28 Management* 2009;63:94-108.
- 29 [11] Cabello J, Fernández N, Alcaraz-Segura D, Oyonarte C, Piñeiro G, Altesor A, Delibes
30 M, Paruelo JM. The Ecosystem Functioning Dimension in Conservation: insights
31 from remote sensing. *Biodiversity and Conservation* 2012;21:3287-3305.
- 32 [12] Fisher B, Turner RK, Morling P. Defining and classifying ecosystem services for deci-
33 sion making. *Ecological Economics*, 2008;643-653.
- 34 [13] (MA) Millennium Ecosystem Assessment. Washington, DC: Island Press 2005.
- 35 [14] Martín-López BE, Gómez-Baggethun M, García-Llorente M, Montes C. Trade-offs
36 across value-domains in ecosystem services assessment. *Ecological Indicators*
37 2014;37:220–228

- 1 [15] Vitousek PM, Mooney HA, Lubchenco J, Melillo JM. Human domination of Earth's
2 ecosystems. *Science* 1997;277(5325), 494
- 3 [16] Alcaraz-Segura D, Paruelo JM, Epstein HE, Cabello J. Environmental and Human
4 Controls of Ecosystem Functional Diversity in Temperate South America. *Remote*
5 *Sensing* 2013;5(1):127-154
- 6 [17] Matthews WJ, Vaughn CC, Gido, KB, Marsh-Matthews E. Southern plains rivers, In:
7 *Rivers of North America*, eds A. Benke, C.E. Cushing. Academic Press 2005.
- 8 [18] Vaughn CC, and M. Pyron Population ecology of the endangered Ouachita Rock
9 Pocketbook mussel, *Arkansia wheeleri* (Bivalvia: Unionidae), in the Kiamichi River,
10 Oklahoma. *American Malacological Bulletin* 1995;11:145-151.
- 11 [19] Vaughn CC, 2000. Changes in the mussel fauna of the Red River drainage: 1910 –
12 present. In: *Proceedings of the First Freshwater Mussel Symposium*, edited by R. A.
13 Tankersley, D. I. Warmolts, G. T. Watters, B. J. Armitage, P. D. Johnson, and R. S.
14 Butler, Ohio Biological Survey, pp. 225-232 Columbus, Ohio.
- 15 [20] Galbraith HS, Spooner DE, Vaughn CC. Status of rare and endangered freshwater
16 mussels in southeastern Oklahoma rivers. *Southwestern Naturalist* 2008;53: 45-50.
- 17 [21] Agbenyega O, Burgess PJ, Cook M, Morris J. Application of an ecosystem function
18 framework to perceptions of community woodlands. *Land Use Policy*
19 2009;26:551-557
- 20 [22] Winkler R. Valuation of ecosystem goods and services, An integrated dynamic ap-
21 proach. *Ecological Economics* 2006;59:82-93
- 22 [23] Nelson E, Sander H, Hawthorne P, Conte M, Ennaanay D. Projecting Global Land-
23 Use Change and Its Effect on Ecosystem Service Provision and Biodiversity with
24 Simple Models. *PLoS ONE* 2010;5(12): e14327. doi:10.1371/journal.pone.0014327
- 25 [24] Eggleston S, Buendia L, Miwa, K, Ngara T, Tanabe K. IPCC Guidelines for National
26 Greenhouse Gas Inventories, Volumen 4. Japan: Institute for Global Environmental
27 Strategies. 2006.
- 28 [25] Potter KN, Derner JD. Soil carbon pools in central Texas: Prairies, restored grass-
29 lands, and croplands. *Soil Water Conservation* 2006;124-128.
- 30 [26] Smith J, Heath L, Skog K, Birdsey R. Methods for calculating forest ecosystem and
31 harvested carbon with standard estimates for forest types of the United States. New-
32 town Square, PA: US Department of Agriculture 2006.
- 33 [27] Goins CR, and Goble D. *Historical Atlas of Oklahoma*, fourth ed., 286 pp., University
34 of Oklahoma Press, Norman. 2006
- 35 [28] Watkins BW. *Reconstructing the Choctaw Nation of Oklahoma, 1894-1898: Land-*
36 *scape and Settlement on the Eve of Allotment*, 202 pp, Oklahoma State University,
37 Stillwater, 2007

- 1 [29] Fry JG, Xian S, Jin J, Dewitz C, Homer L, Yang C, Barnes N, and J. Wickham. Com-
2 pletion of the 2006 National Land Cover Database for the Conterminous United
3 States. *PE&RS* 2011;858-864.
- 4 [30] Villa F, Bagstad KJ, Voigt B, Johnson GW, Portela R, Honzak M, Batker D. A method-
5 ology for adaptable and robust ecosystem services assessment. *PLoS ONE*
6 2014;9(3):e91001.
- 7 [31] Jackson B, Timothy P, Sinclair F. Polyscape: A GIS mapping framework providing
8 efficient and spatially explicit landscape-scale valuation of multiple ecosystem serv-
9 ices. *Landscape and Urban Planning* 2013;112:74–88
- 10 [32] Tran TV. Measuring land cover change at high spatio-temporal scales, 95 pp, Univer-
11 sity of Oklahoma; 2013.
- 12 [33] Sleeter BM, Sohl TL, Loveland TR, Auch RF, Acevedo W, Drummond MA, Sayler
13 KL, and Stehman SV. Land-cover change in the conterminous United States from
14 1973 to 2000, *Global Environmental Change*, 2013;23(4), 733-748
- 15 [34] Reker RR, Sayler, KL, Friesz AM, Sohl TL, Bouchard MA, Sleeter BM, Sleeter RR,
16 Wilson TS, Griffith GE, and Knuppe ML. Mapping and modeling of land use and
17 land cover in the eastern United States from 1992 through 2050, in *Baseline and pro-*
18 *jected future carbon storage and greenhouse gas fluxes in ecosystems of the eastern*
19 *United States: U.S. Geological Survey Professional Paper 1804*, edited by Z. Zhu and
20 B. C. Reed, 2014;pp. 27-54.
- 21 [35] Jawarneh RN, and Julian JP. Development of an accurate fine-resolution land cover
22 timeline: Little Rock, Arkansas, USA (1857–2006). *Applied Geography* 2012;35(1–
23 2)104-113.

